# Exhibit 24

Memorandum from Stephen D. Page, Director, Office of Air Quality Planning Standards, EPA, Re: Modeling Procedures for Demonstrating Compliance with PM2.5 NAAQS (March 23, 2010)



# UNITED STATES ENVIRONMENTAL PROTECTION AGENCY RESEARCH TRIANGLE PARK, NC 27711

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OFFICE OF AIR QUALITY PLANNING AND STANDARDS

# MEMORANDUM

SUBJECT: Modeling Procedures for Demonstrating-Compliance with PM2.5 NAAQS

FROM: Stephen D. Page, Director Musher Office of Air Quality Planning and Standards

# TO: See Addressees

This memorandum addresses the need for recommendations regarding appropriate dispersion modeling procedures which can be used to demonstrate compliance with  $PM_{2.5}$  National Ambient Air Quality Standards (NAAQS). The need for these recommendations arises from several recent regulatory actions and proposals which increase the likelihood that applicants for permits under the new source review (NSR) and prevention of significant deterioration (PSD) programs may be required to demonstrate compliance with  $PM_{2.5}$  NAAQS rather than relying upon the  $PM_{10}$  surrogate policy established in 1997. These recommendations are intended to facilitate appropriate and consistent implementation of current guidance regarding  $PM_{2.5}$  dispersion modeling contained in the *Guideline on Air Quality Models*, Appendix W to 50 CFR Part 51, while acknowledging that such guidance is somewhat limited in detail due to technical issues associated with  $PM_{2.5}$  modeling.

This memorandum provides recommendations on two aspects of the modeling procedures for demonstrating compliance with the  $PM_{2.5}$  NAAQS. First, this memorandum discusses some of the technical issues that must be addressed by any applicant or permitting authority that is seeking to rely on the  $PM_{10}$  surrogate policy. Second, this memorandum provides additional information on modeling procedures to demonstrate compliance with  $PM_{2.5}$  NAAQS without relying upon the  $PM_{10}$  surrogate policy.

# BACKGROUND

On July 18, 1997, EPA revised the NAAQS for particulate matter to add new annual and 24-hour standards for fine particles using  $PM_{2.5}$  as the indicator. EPA revised the 24-hour NAAQS for  $PM_{2.5}$  on September 21, 2006, reducing the standard from 65 µg/m<sup>3</sup> to 35 µg/m<sup>3</sup>. EPA also retained the previous 1997 annual standard for  $PM_{2.5}$  and the 24-hour standard for  $PM_{10}$ , while revoking the previous annual standard for  $PM_{10}$ . For attainment of the new 24-hour  $PM_{2.5}$  NAAQS based on ambient monitoring, the average of the 98<sup>th</sup> percentile 24-hour values

over three years of monitoring must not exceed 35  $\mu$ g/m<sup>3</sup>. The annual PM<sub>2.5</sub> NAAQS is set at 15  $\mu$ g/m<sup>3</sup> based on the average of the annual mean PM<sub>2.5</sub> concentrations over three years.

Citing significant technical difficulties with respect to PM2.5 monitoring, emissions estimation, and modeling, EPA established a policy, known as the PM<sub>10</sub> surrogate policy, on October 23, 1997. This policy allowed permit applicants to use compliance with the applicable  $PM_{10}$  requirements as a surrogate approach for meeting  $PM_{2.5}$  NSR requirements until the technical difficulties were resolved. On May 16, 2008, EPA promulgated final rules governing the implementation of the NSR program for PM2.5, which included a "grandfathering provision" allowing applicants for federal PSD permits covered by 40 CFR § 52.21, with complete permit applications submitted as of July 15, 2008, to continue relying on the  $PM_{10}$  surrogate policy. In response to a petition challenging the continued use of the  $PM_{10}$  surrogate policy for issuing PSD permits, on June 1, 2009, EPA issued a 3-month administrative stay of the grandfathering provision for PM<sub>2.5</sub> affecting federal PSD permits to give EPA time to propose repealing the challenged grandfathering provision. On September 16, 2009, the original 3-month stay was extended to June 22, 2010, to allow additional time for EPA to formally propose repeal of the grandfathering provision from the PM<sub>2.5</sub> NSR implementation rule for federal PSD permits issues under 40 CFR § 52.21. On February 11, 2010, EPA published its proposal to repeal the grandfathering provision in the Federal Register at 75 FR 6827. These actions cite the fact that the technical difficulties which necessitated the  $PM_{10}$  surrogate policy have been largely, although not entirely, resolved.

As part of the proposed rulemaking to repeal the grandfathering provision contained in the federal PSD program, EPA has also proposed to end the use of the PM<sub>10</sub> surrogate policy for state PSD programs that EPA has approved as part of the state implementation plan (SIP) under 40 CFR § 51.166. Under the PSD programs for PM<sub>2.5</sub> currently in effect for SIP-approved states, states would be allowed to continue using the PM<sub>10</sub> surrogate policy until May 2011, or until EPA approves the revised SIP for PM<sub>2.5</sub>, whichever occurs first. While we continue to allow states to use the PM<sub>10</sub> surrogate policy during their transition to the new PM<sub>2.5</sub> requirements, we have also made it clear that the policy needs to be implemented by taking into account court decisions that address the surrogacy concept. Accordingly, an applicant seeking a PSD permit under a SIP-approved PSD program may still rely upon the PM<sub>10</sub> surrogate policy as long as (1) the appropriateness of the PM<sub>10</sub>-based assessment for determining PM<sub>2.5</sub> compliance has been adequately demonstrated based on the specifics of the project; and (2) the applicant can show that a PM<sub>2.5</sub> analysis is not technically feasible. Absent such demonstrations, applicants would be required to submit a PM<sub>2.5</sub>-based assessment to demonstrate compliance with the PM<sub>2.5</sub> standards, in addition to meeting the other requirements under the NSR/PSD programs.

#### **PM<sub>10</sub> SURROGACY DEMONSTRATIONS**

Given the need for applicants that continue to rely on the PM<sub>10</sub> surrogate policy to demonstrate the appropriateness of the policy based on the specifics of the project, we feel that it is appropriate and timely to address some of the technical issues associated with a surrogacy demonstration. EPA's August 12, 2009, Administrative Order in response to petitions regarding the Title V permit for Louisville Gas and Electric Company (LG&E), Trimble Generating Station, provides a brief summary of the case law history that bears on the PM<sub>10</sub> surrogacy issue which suggests that an appropriateness demonstration "would need to address the differences between  $PM_{10}$  and  $PM_{2.5}$ ."<sup>1</sup> The LG&E order cites two examples in this regard: 1) "emission controls used to capture coarse particles may be less effective in controlling  $PM_{2.5}$ "; and 2) "particles that make up  $PM_{2.5}$  may be transported over long distances while coarse particles normally only travel short distances." These examples serve to highlight the two main aspects of PSD permitting for which the appropriateness of the surrogate policy should be demonstrated: 1) the Best Available Control Technology (BACT) emission control technology assessment; and 2) the ambient air quality impact assessment to demonstrate compliance with the applicable NAAQS.

While acknowledging "an evolving understanding of the technical and legal issues associated with the use of the  $PM_{10}$  Surrogate Policy," the LG&E order offers two steps as possible approaches for making an appropriateness demonstration, without suggesting that the "two steps are necessary or sufficient to demonstrate that  $PM_{10}$  is a reasonable surrogate for  $PM_{2.5}$ " and clearly stating that "these two steps are not intended to be the exclusive list of possible demonstrations" regarding surrogacy. The two steps offered in the LG&E order are primarily relevant to the appropriateness demonstration regarding emission controls under BACT, while the discussion here will be focused on the appropriateness demonstration in relation to ambient air impacts.

Given the range of application-specific factors that may need to be addressed for an appropriateness demonstration in relation to ambient air impacts, it is not practical to provide detailed guidance regarding how to conduct such demonstrations. However, the following list identifies some of the "differences between  $PM_{10}$  and  $PM_{2.5}$ " in relation to ambient air impacts that should be addressed in the development of a surrogacy demonstration:

- 1. While EPA revoked in 2006 the annual PM<sub>10</sub> standard that was in effect when the surrogate policy, the surrogacy demonstration would still need to address the appropriateness of the PM<sub>10</sub> surrogate policy in relation to the annual PM<sub>2.5</sub> standard, and would likely require a modeling analysis of annual PM<sub>10</sub> impacts.
- 2. The current 24-hour NAAQS of 35  $\mu$ g/m<sup>3</sup> is well below the previous level of 65  $\mu$ g/m<sup>3</sup> that was in effect when the PM<sub>10</sub> surrogate policy was established. The background monitored levels of PM<sub>2.5</sub> are, therefore, likely to account for a more significant fraction of the cumulative impacts from a modeling analysis relative to the current 24-hour PM<sub>2.5</sub> NAAQS than for PM<sub>10</sub>.
- 3. Secondary formation of PM<sub>2.5</sub> from emissions of NO<sub>x</sub>, SO<sub>x</sub> and other compounds from sources across a large domain will often contribute significantly to the total ambient levels of PM<sub>2.5</sub>, and may be the dominant source of ambient PM<sub>2.5</sub> in some cases. In contrast, secondarily formed particles are less likely to be significant portion of PM<sub>10</sub>, which may result in significant differences in the spatial and temporal patterns of ambient impacts between PM<sub>2.5</sub> and PM<sub>10</sub>.

<sup>&</sup>lt;sup>1</sup> A discussion of the case law that bears on the  $PM_{10}$  surrogacy issue also appears in the February 11, 2010, proposed rule at 75 FR 6831-6832.

4. The probabilistic form of the PM<sub>2.5</sub> NAAQS, based on the multiyear average of the 98<sup>th</sup> percentile for the daily standard, differs from the expected exceedance form of the PM<sub>10</sub> NAAQS, which allows the standard to be exceeded once per year on average using the high-sixth-high (H6H) value over 5 years. These differences affect the temporal and spatial characteristics of the ambient air impacts of PM<sub>10</sub> and PM<sub>2.5</sub>. Differences in the form of the NAAQS also complicate the process of combining modeled impacts with monitored background levels to estimate cumulative impacts under the NSR/PSD permitting programs, as well as the determination of whether modeled impacts from the facility will cause a significant contribution to any modeled violations of the NAAQS that may occur.

These factors complicate the viability of demonstrating the appropriateness of the  $PM_{10}$  surrogate policy to comply with the requirement for a  $PM_{2.5}$  ambient air quality impact assessment. In light of these complications, applicants may elect to use  $PM_{2.5}$  dispersion modeling to explicitly meet the requirement of an ambient air quality impact assessment under the PSD permitting program, provided that the technical difficulties with respect to  $PM_{2.5}$  monitoring, emissions estimation, and modeling have been sufficiently resolved in relation to the specific application.

For surrogacy demonstrations, it is assumed that as an initial step the applicant will have conducted an appropriate dispersion modeling analysis which demonstrates compliance with the  $PM_{10}$  NAAQS, including an analysis of annual  $PM_{10}$  impacts to address item 1. A simple example illustrating when a  $PM_{10}$  modeling analysis might serve as a surrogate for  $PM_{2.5}$  modeling would be if a clearly conservative assumption is made that all  $PM_{10}$  emissions are  $PM_{2.5}$ , and the modeled  $PM_{10}$  impacts are taken as a direct surrogate for  $PM_{2.5}$  impacts and compared to the  $PM_{2.5}$  NAAQS. If an adequate accounting for contributions from background  $PM_{2.5}$  concentrations to the cumulative impact assessment can be made, and a reasonable demonstration that the modeled  $PM_{10}$  emission inventory adequately accounted for potential nearby sources of  $PM_{2.5}$ , then the appropriateness of surrogacy could be reasonably found in this example. An analysis of source-specific  $PM_{2.5}/PM_{10}$  emission factor ratios may also support the assumption of a more realistic, yet still conservative approach for taking a ratio of modeled  $PM_{10}$  ambient impacts to provide conservative estimates of  $PM_{2.5}$  impacts.

While additional modeling analyses, short of explicit  $PM_{2.5}$  modeling, may also be used to the support the surrogacy demonstration in some cases, it is important to make a clear distinction between modeling analyses for purposes of surrogacy demonstrations and modeling analyses that are intended to explicitly demonstrate compliance with the  $PM_{2.5}$  standards. The distinction between these two types of modeling analyses may not always be clear, but one important distinction is whether or not a  $PM_{2.5}$  emission inventory has been developed as the basis for the modeling. The distinction between these types of modeling is important because modeling procedures that may be considered appropriate for one type of analysis may not be appropriate for the other. The following section elaborates further on this point.

#### PM<sub>2.5</sub> MODELING ANALYSES

The differences between  $PM_{10}$  and  $PM_{2.5}$  described above in relation to surrogacy demonstrations, especially items 2 through 4, also have implications on how best to conduct an explicit  $PM_{2.5}$  NAAQS compliance demonstration through dispersion modeling. Due to the potentially significant contribution from secondary formation of  $PM_{2.5}$ , and the more prominent role of monitored background concentrations of  $PM_{2.5}$  in the cumulative analysis, certain aspects of standard modeling practices used for  $PM_{10}$  and other criteria pollutants may not be appropriate for  $PM_{2.5}$ . Our recommendations for addressing these issues in terms of explicit  $PM_{2.5}$  modeling analyses are described in more detail below.

Given the issues listed above, and especially the important contribution from secondary formation of  $PM_{2.5}$ , which is not explicitly accounted for by the dispersion model, PSD modeling of  $PM_{2.5}$  should currently be viewed as screening-level analyses, analogous to the screening nature of the guidance in Section 5.2.4 of Appendix W regarding dispersion modeling for NO<sub>2</sub> impacts given the importance of chemistry in the conversion of NO emissions to ambient NO<sub>2</sub>. The screening recommendations presented below for demonstrating compliance with the  $PM_{2.5}$  NAAQS through dispersion modeling have been developed with the factors listed above in mind. As with any modeling analysis conducted under Appendix W, alternative models and methods may be considered on a case-by-case basis, subject to approval by the Regional Office in accordance with the recommendations in Section 3.2 on "Use of Alternative Models."

The following sections describe the recommended modeling methods for the two main stages in a typical PSD ambient air quality analysis: 1) preliminary significant impact analysis; and 2) cumulative impact assessment. The rationale for the recommendations is also provided.

## **Preliminary Significant Impact Analysis**

The initial step in air quality impact assessments under NSR/PSD is typically a significant impact level analysis to determine whether the proposed emissions increase from the proposed new or modified source (i.e., project emissions) would have a "significant" ambient impact. Thus, the first step of the ambient impact analysis is to determine whether those emissions would result in ambient air concentrations that exceed a de minimis level, referred to as the Significant Impact Level (SIL). If modeled impacts from the facility do not exceed the SIL, then the permitting authority may be able to conclude, based on this preliminary analysis, that the project would not cause or contribute to a violation of the NAAQS. Under these circumstances, EPA would not consider it necessary for the facility to conduct a more comprehensive cumulative impact assessment that would involve modeling the facility's total emissions along with emissions from other nearby background sources, and combining impacts from the modeled emission inventory with representative ambient monitored background concentrations to estimate the cumulative impact levels for comparison to the NAAOS. The SIL is also used to establish the significant impact area of the facility for purposes of determining the geographic range of the background source emission inventory that would be appropriate should a cumulative impact assessment be necessary.

EPA's 2007 proposed rule to establish PSD increments, SILs, and a Significant Monitoring Concentration (SMC) for  $PM_{2.5}$  included three options for the  $PM_{2.5}$  SILs for both the 24-hour and annual NAAQS. Until the  $PM_{2.5}$  SILs are finalized, the proposed SILs may not be presumed to be appropriate de minimis impact levels. However, EPA does not preclude states from adopting interim de minimis impact levels for  $PM_{2.5}$  to determine whether a cumulative impact analysis will be necessary, provided that states prepare an appropriate record to support the value used. Such de minimis levels do not necessarily have to match any of the SILs that have been proposed for  $PM_{2.5}$ , but the levels proposed by EPA and the record supporting EPA's proposed rule could be considered in the state's determination.

The modeling methods used in this initial significant impact assessment phase of the PM<sub>2.5</sub> analysis, based on either a state's interim de minimis levels or EPA-finalized SILs, are similar to the methods used for other pollutants, including the use of maximum allowable emissions. However, due to the probabilistic form of the NAAQS, we recommend that the highest average of the modeled annual averages across 5 years for National Weather Service (NWS) meteorological data or the highest modeled annual average for one year of site-specific meteorological data be compared to the annual screening level (SIL). Similarly, the highest average of the maximum 24-hour averages across 5 years for NWS meteorological data or the highest modeled 24-hour average for one year of site-specific meteorological data should be compared to the 24-hour screening level (SIL).

Using the average of the highest values across the years modeled preserves one aspect of the form of the NAAQS, while using the average of the <u>first</u> highest 24-hour averages rather than the 98<sup>th</sup> percentile (8<sup>th</sup> highest) values from the distribution is consistent with the screening-level nature of the analysis. In addition, since the PM<sub>2.5</sub> NAAQS is based on air quality levels averaged over time, it is appropriate to use an average modeled impact for comparison to the SIL since that will more accurately characterize the modeled contribution from the facility in relation to the NAAQS than use of the highest modeled impacts from individual years. At the present time, the dispersion modeling recommendations presented here are based on modeling only the primary or direct PM<sub>2.5</sub> emissions from the facility.

#### **Cumulative Impact Assessment**

Unless modeled ambient air concentrations of  $PM_{2.5}$  from the project emissions are shown to fall below the state's de minimis level or EPA's promulgated SIL (when finalized), then a cumulative impact assessment would be necessary to account for the combined impact of facility emissions, emissions from other nearby sources, and representative background levels of  $PM_{2.5}$  within the modeling domain. The cumulative impacts are then compared to the NAAQS to determine whether the facility emissions will cause or contribute to a violation of the NAAQS. Several aspects of the cumulative impact assessment for  $PM_{2.5}$  will be comparable to assessments conducted for other criteria pollutants, while other aspects will differ due to the issues identified above.

## Modeling Inventory

The current guidance on modeling emission inventories contained in Section 8.1 of Appendix W will generally be applicable for the  $PM_{2.5}$  modeling inventory, recognizing that these recommendations only address modeling of primary  $PM_{2.5}$  emissions. The guidance in Appendix W addresses the appropriate emission level to be modeled, which in most cases is the maximum allowable emission rate under the proposed permit. Nearby sources that are expected to cause a significant concentration gradient in the vicinity of the facility should generally be included in the modeled inventory. Since modeling of  $PM_{2.5}$  emissions has not been a routine requirement to date, the availability of an adequate  $PM_{2.5}$  emission inventory for background sources may not exist in all cases. Recommendations for developing  $PM_{2.5}$  emission inventories for use in PSD applications will be addressed separately, but existing  $PM_{10}$  inventories may provide a useful starting point for this effort.

#### Monitored Background

The determination of representative background monitored concentrations of  $PM_{2.5}$  to include in the  $PM_{2.5}$  cumulative impact assessment will entail different considerations from those for other criteria pollutants. An important aspect of the monitored background concentration for  $PM_{2.5}$  is that the monitored data should account for the contribution of secondary  $PM_{2.5}$  formation representative of the modeling domain. As with other criteria pollutants, consideration should also be given to the potential for some double-counting of the impacts from modeled emissions that may be reflected in the background monitoring, but this should generally be of less importance for  $PM_{2.5}$  than the representativeness of the monitor for secondary contributions. Also, due to the important role of secondary  $PM_{2.5}$ , background monitored concentrations of  $PM_{2.5}$  are likely to be more homogeneous across the modeling domain in most cases, compared to other pollutants. We plan to address separately more detailed guidance on the determination of representative background concentrations for  $PM_{2.5}$ .

## Comparison to NAAQS

Combining the modeled and monitored concentrations of  $PM_{2.5}$  for comparison to the  $PM_{2.5}$  NAAQS also entails considerations that differ from those for other criteria pollutants, due to the issues identified above. Given the importance of secondary contributions for  $PM_{2.5}$  and the typically high background levels relative to the NAAQS for  $PM_{2.5}$ , greater emphasis is placed on the monitored background contribution relative to the modeled inventory. Also, given the probabilistic form of the  $PM_{2.5}$  NAAQS, careful consideration must be given to how the monitored and modeled concentrations are combined to estimate the cumulative impact levels.

The representative monitored  $PM_{2.5}$  design value, rather than the overall maximum monitored background concentration, should be used as a component of the cumulative analysis. The  $PM_{2.5}$  design value for the annual averaging period is based on the 3-year average of the annual average  $PM_{2.5}$  concentrations; for the 24-hour averaging period, the design value is based on the 3-year average of the 98<sup>th</sup> percentile 24-hour average  $PM_{2.5}$  concentrations for the daily standard. Details regarding the determination of the 98<sup>th</sup> percentile monitored 24-hour value

based on the number of days sampled during the year are provided in the ambient monitoring regulations, Appendix N to 40 CFR Part 50.

The modeled annual concentrations of (primary)  $PM_{2.5}$  to be added to the monitored annual design value should be computed using the same procedure used for the initial significant impact analysis based on the highest average of the modeled annual averages across 5 years for NWS meteorological data or the highest modeled annual average for one year of site-specific meteorological data. The resulting cumulative annual concentration would then be compared to the annual  $PM_{2.5}$  NAAQS of 15  $\mu$ g/m<sup>3</sup>.

For the 24-hour NAAQS analysis, the modeled concentrations to be added to the monitored 24-hour design value should be computed using the same procedure used for the preliminary analysis based on the highest average of the maximum modeled 24-hour averages across 5 years for NWS meteorological data or the maximum modeled 24-hour average for one year of site-specific meteorological data. As noted above, use of the average modeled concentration across the appropriate time period more accurately characterizes the modeled impact from individual years, while using the average of the first highest 24-hour averages rather than the 98<sup>th</sup> percentile (8<sup>th</sup> highest) values is consistent with the screening nature of PM<sub>2.5</sub> dispersion modeling. Furthermore, combining the 98<sup>th</sup> percentile monitored with the 98<sup>th</sup> percentile modeled concentrations for a cumulative impact assessment could result in a value that is below the 98<sup>th</sup> percentile of the combined cumulative distribution and would, therefore, not be protective of the NAAQS.

The recommendations provided above constitute a First Tier modeling analysis for  $PM_{2.5}$  compliance demonstrations. For applications where impacts from primary  $PM_{2.5}$  emissions are not temporally correlated with background  $PM_{2.5}$  levels, combining the modeled and monitored contributions as described above may be overly conservative. In these cases, a Second Tier modeling analysis may be considered that would involve combining the monitored and modeled  $PM_{2.5}$  concentrations on a seasonal or quarterly basis, and re-sorting the total impacts across the year to determine the cumulative design value. We plan to provide separately additional details regarding this Second Tier, including a discussion of circumstances where this approach may be appropriate.

#### Determining Significant Contributions to Modeled Violations

If the cumulative impact assessment following these screening recommendations results in modeled violations of the  $PM_{2.5}$  NAAQS, then the applicant will need to determine whether the facility emissions are causing a significant contribution to those modeled violations. A "significant contribution" determination is based on a comparison of the modeled impacts from the project emissions associated with the modeled violation to the appropriate SIL. The significant contribution determination should be made following the same procedures used during the initial significant impact analysis, based on a comparison of the average of the modeled concentrations at the receptor location showing the violation, across 5 years for NWS meteorological data and the highest modeled concentration for one year of site-specific meteorological data. For a violation of the annual NAAQS, the average of the annual values at the affected receptor(s) is compared to the SIL, while the average of the highest 24-hour average concentrations at the affected receptor(s) should be used for the 24-hour NAAQS. Use of the average modeled concentration is appropriate in this context since it is consistent with the actual contribution of the facility to the cumulative impacts at the receptor(s) showing violations and accounts for the fact that modeled violations of the 24-hour NAAQS represent average impacts across the modeling period.

# **Synopsis**

Significant Impact Analysis: Compare the average of the highest modeled individual year's annual averages and the average of the first highest individual year's 24-hour average concentrations from project emissions to their respective screening levels, which may be based on the state's de minimis levels or EPA-finalized SILs. If modeled impacts exceed the screening levels, a cumulative impact assessment would need to be performed.

<u>Cumulative Impact Assessment:</u> Develop an emission inventory of background sources to be included in the modeling analysis using traditional guidance. That would include using the significant impact area established in the initial significant impact analysis, plus a 50-km annular ring to determine the geographic extent of the background emission inventory. From data obtained within this combined area, compare the average of the highest modeled individual year's annual averages and the average of the first highest individual year's 24-hour averages, plus representative background monitored concentrations, to their respective NAAQS. Monitored background concentrations are based on the 3-year average of the annual PM<sub>2.5</sub> concentrations, and the 3-year average of the 98<sup>th</sup> percentile 24-hour averages. To determine whether the proposed project's emissions cause a significant contribution to any modeled violations of the NAAQS, the proposed project's impacts at the affected receptor(s) are determined based on the average of the highest modeled individual years' annual averages and average of the highest modeled individual years' annual averages and average of the highest modeled individual years' annual averages and average of the first highest individual years' annual averages and average of the first highest individual years' annual averages and average of the first highest individual years' 24-hour averages from the proposed project's emissions, and are compared to the state's de minimis levels or EPA-finalized SILs.

# **Additional Caveats**

A few additional caveats should be considered while implementing these recommendations:

 The current preferred dispersion model for near-field PM<sub>2.5</sub> modeling, AERMOD, does not account for secondary formation of PM<sub>2.5</sub>. Therefore, any secondary contribution of the facility's or other modeled source's emissions is not explicitly accounted for. While representative background monitoring data for PM<sub>2.5</sub> should adequately account for secondary contribution from background sources in most cases, if the facility emits significant quantities of PM<sub>2.5</sub> precursors, some assessment of their potential contribution to cumulative impacts as secondary PM<sub>2.5</sub> may be necessary. In determining whether such contributions may be important, keep in mind that peak impacts due to facility primary and secondary PM<sub>2.5</sub> are not likely to be well-correlated in space or time, and these relationships may vary for different precursors. We plan to issue separately additional guidance regarding this issue.

- 2. While dry and/or wet deposition may be important processes when estimating ambient concentrations of particulate matter (PM) in general, these factors are expected to be minor for PM<sub>2.5</sub> due to the small particle size. In addition, there may be additional uncertainty associated with deposition modeling for PM<sub>2.5</sub> due to the variable makeup of the constituent elements for PM<sub>2.5</sub> and the fact that deposition properties may vary depending on the constituent elements of PM<sub>2.5</sub>. Therefore, use of deposition algorithms to account for depletion in estimating ambient PM<sub>2.5</sub> concentrations should be done with caution and only when clear documentation and justification of the deposition parameters is provided.
- 3. While EPA has proposed PSD increments for PM<sub>2.5</sub>, the increments have not been finalized yet. Until the increments are finalized, no increment analysis is required for PM<sub>2.5</sub>. However, it should be noted that some of the recommendations presented here in relation to NAAQS modeling analyses may need to be modified for PM<sub>2.5</sub> increment analyses due to the differences between the forms of the NAAQS and increments. We plan to provide further clarification of these differences separately, once the increments are finalized.

This memorandum presents EPA's views on these issues concerning modeling procedures for demonstrating compliance with the  $PM_{2.5}$  NAAQS. The statements in this memorandum do not bind State and local governments and the public as a matter of law. If you have any questions concerning this memorandum, please contact Tyler Fox, Leader, Air Quality Modeling Group at (919) 541-5562.

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